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STUDIES

Increased ploidy of Butomus umbellatus in introduced populations is not associated with higher phenotypic plasticity to N and P

Nathan E. Harms^{1,2,*}, James T. Cronin² and John F. Gaskin³

¹U.S. Army Engineer Research and Development Center, Aquatic Ecology and Invasive Species Branch, 3909 Halls Ferry Road, Vicksburg, MS 39180, USA, ²Department of Biological Sciences, Louisiana State University, 202 Life Sciences Building, Baton Rouge, LA 70803, USA, ³U.S. Department of Agriculture, Agricultural Research Service, 1500 N. Central Avenue, Sidney, MT 59270, USA

*Corresponding author's e-mail address: Nathan.E.Harms@usace.army.mil

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Abstract

Separate introductions or post-introduction evolution may lead to multiple invader genotypes or cytotypes that differ in growth rates, biomass or chemical profile responses (phenotype) to a range of environments. If the invader has high trait plasticity to a range of resource levels, then sediment N or P enrichment may enhance invasiveness. However, the ways in which ploidy, plasticity, and available N or P interact are unknown for most species despite the potential to explain spread and impacts by invaders with multiple introduced lineages. We conducted a common garden experiment with four triploid and six diploid populations of Butomus umbellatus, collected from across its invasive range in the USA. Plants were grown under different N or P nutrient levels (4, 40, 200, 400 mg L-1 N; 0.4, 4, 40 mg L-1 P) and we measured reaction norms for biomass, clonal reproduction and tissue chemistry. Contrary to our expectation, triploid B. umbellatus plants were less plastic to variation in N or P than diploid B. umbellatus in most measured traits. Diploid plants produced 172 % more reproductive biomass and 57 % more total biomass across levels of N, and 158 % more reproductive biomass and 33 % more total biomass across P than triploid plants. Triploid plants had lower shoot:root ratios and produced 30 % and 150 % more root biomass than diploid plants in response to increases in N and P, respectively. Tissue chemistry differed between cytotypes but plasticity was similar; N was 8 % higher and C:N ratio was 30 % lower in triploid than diploid plants across levels of N and plant parts, and N was 22 % higher and C:N ratio 27 % lower across levels of P and plant parts. Our results highlight differences in nutrient response between cytotypes of a widespread invader, and we call for additional field studies to better understand the interaction of nutrients and ploidy during invasion.

Keywords: Biomass allocation; flowering rush (Butomus umbellatus); nutrient enrichment; phenotypic plasticity; plant invasion; polyploidy.

Introduction

Phenotypic responses to changing environmental conditions (i.e. phenotypic plasticity) explain the success of some invasive species during establishment and their subsequent spread to new areas (Davidson *et al.* 2011; Higgins and Richardson 2014; Turner et al. 2015). The ability to colonize and establish under varying conditions and in a range of environments can be highly advantageous for an invading species (Richards et al. 2005). This may be particularly important if the phenotypic response leads

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to a greater fitness advantage for the invader than the response of species in recipient communities (Leishman and Thomson 2005; Richardson and Pyšek 2006). Plasticity may also evolve in response to new environments encountered after introduction, for example along an invasion front (Chevin and Lande 2011), or through generation of novel genotypes as a result of genetic crossings between invaders originating from different locations (i.e. hybrid vigor; Williams et al. 2005; Lavergne and Molofsky 2007; Xu et al. 2010). The importance of phenotypic plasticity for defining niche breadth, especially with regards to successful plant invasions, has been well studied for more than a decade, with research focused on variation in trait plasticity to salinity (Richards et al. 2005; Meyerson et al. 2020), water availability (Leishman and Thomson 2005), competition (Peperkorn et al. 2005; Peacor et al. 2006), light (DeWalt et al. 2004; Siebenkäs et al. 2015), shade (Griffith and Sultan 2005), herbivory (Simoes and Baruch 1991; Schierenbeck et al. 1994; Agrawal et al. 2002; Ashton and Lerdau 2008; Bhattarai et al. 2017) and nutrients (Burns 2004; Hastwell and Panetta 2005; Leishman and Thomson 2005).

Nutrient enrichment, for example resulting from fertilizer use in agricultural areas, and its role in promoting plant invasions has been examined for a number of terrestrial (e.g. Huenneke et al. 1990; Lowe et al. 2002; Gross et al. 2005; Vasquez et al. 2008) and aquatic systems (e.g. Fan et al. 2013; Yu et al. 2018). Studies that examine plant phenotypic (e.g. growth rate, biomass allocation, tissue chemistry) variation in response to nutrient conditions (elevated or reduced) between congeneric species have led to a better mechanistic understanding of some plant invasions (Woo and Zedler 2002; Kolb and Alpert 2003; Godoy et al. 2011; Funk 2013) and the conditions under which some species, but not others, become invasive (Flanagan et al. 2015). In fact, understanding the role of trait plasticity in plant response to available nutrients (e.g. N, P) may be key to predicting areas at risk of invasion (Wang et al. 2017), assessing the performance of introduced species (Seabloom et al. 2015), managing invasive species (Blumenthal 2006) and understanding trophic interactions involving the invader (e.g. Room et al. 1989; Steinger and Müller-Schärer 1992; Center and Dray 2010). Although a number of traits have been associated with successful invaders, plasticity in nitrogen use efficiency, photosynthetic rate, biomass production and allocation (e.g. shoot:root ratio) and tissue chemistry (e.g. N, C:N) are consistently higher in invasive than non-invasive species across studies (Davidson et al. 2011). In fact, success of invaders in enriched systems may be due in part to their proportionally greater phenotypic response across a range of nutrient levels in comparison to native species (i.e. 'jack-of-alltrades' sensu Richards et al. 2006) (Huenneke et al. 1990; Lowe et al. 2002; Gross et al. 2005; Vasquez et al. 2008), their greater response to nutrient enhancement ('master-of-some') or interactions between nutrients and herbivory (Wright et al. 2014).

Evolutionary processes occurring during plant invasions can produce genetic structure and generate meaningful differences in invasiveness or response to management activities among and within populations (Bossdorf *et al.* 2005; Sax *et al.* 2005; Ward *et al.* 2008; Gaskin *et al.* 2013). The number and source of introductions, founder or bottleneck effects, hybridization and post-introduction adaptation (e.g. along latitudinal or productivity gradients) may all act to produce populations that diverge in key invasive traits such as reproductive mode or phenology (Kliber *et al.* 2005; Golani *et al.* 2007; Ward *et al.* 2008; Williams *et al.* 2014). For example, the aquatic plant Alternanthera philoxeroides has been introduced into the USA at least twice (Kay and Haller 1982), with introduced biotypes displaying differential susceptibility to herbicides (Kay 1992) and biological control agents (Pan et al. 2012), and responses to nutrients (N. E. Harms, unpubl. data). Another important aspect of genetic variability in invading populations concerns the relationship between chromosome number (ploidy), plasticity to environmental conditions (i.e. niche breadth) and invasiveness (Hahn et al. 2012). The role of increased ploidy in generating generalist phenotypes has been suggested a number of times (Levin 1983; te Beest et al. 2012; Hao et al. 2013), although there are an increasing number of cases where polyploids of invasive taxa tend to have reduced plasticity in important traits related to invasion success. For example, the invasive aquatic plant congeners Ludwigia hexapetala and L. peploides vary in ploidy (L. hexapetala is decaploid and L. peploides is diploid) and are invasive in the USA, but the colonizing form of the diploid L. peploides outperforms L. hexapetala at high nutrient levels, an important characteristic to consider for management where eutrophication occurs (Grewell et al. 2016). The success of invaders with multiple ploidy levels may also depend largely on interactions between ploidy, genome size and environmental variation (Meyerson et al. 2016, 2020). Despite considerable interest in evolutionary processes occurring during plant invasions, and the potential management benefit of better understanding relationships between invasion genetics and phenotypic plasticity to environmental variation, the interplay between genetic variation within invading populations and nutrient response and allocation has not been sufficiently studied (but see Riis et al. 2010; Kettenring et al. 2011; Hahn et al. 2012; Gioria and Osborne 2014).

We conducted a common garden experiment to examine genetic differences in growth, reproduction and tissue chemistry of the invasive wetland plant, Butomus umbellatus, grown hydroponically in increasing concentrations of either N or P. Although the relationship between plasticity of specific traits and fitness remains difficult to demonstrate without long-term demographic studies (Davidson et al. 2011), for clonally reproducing invaders such as B. umbellatus, biomass and reproduction (i.e. generation of vegetative propagules) are often used as fitness proxies (Pan and Price 2001; Eller and Brix 2012; Younginger et al. 2017). Literature reports suggest that triploid B. umbellatus plants have higher fitness under elevated nutrient conditions than diploid plants (Hroudová and Zákravský 1993; Madsen et al. 2016). However, we know little about biomass accumulation and allocation, clonal reproduction, tissue chemistry or plasticity differences in those traits between cytotypes. Therefore, the objectives of this study were to determine whether triploid B. umbellatus plants have higher performance than diploid plants across two nutrient (N and P) gradients and whether triploid plants have higher plasticity of measured traits to increasing nutrients. Specifically, we tested our predictions that (i) triploid plants would produce greater biomass (total, reproductive) than diploid plants across increasing N or P gradients overall, (ii) triploid plants would display increased plasticity to N or P and (iii) tissue chemistry (C, N, P, C:N, C:P, N:P) would similarly differ between cytotypes (diploid, triploid) in response to available nutrients.

Materials and Methods

Study system

Butomus umbellatus is a perennial Eurasian wetland plant species, with diploid and triploid cytotypes, introduced into North America multiple times during the last 100 years from different source areas (Anderson et al. 1974; Kliber et al. 2005). Reproduction in B. umbellatus is primarily clonal (diploid and triploid plants), but can be sexual (diploid plants). Although viable seeds have been reported, they are not as important for spread in the USA as clonal propagules such as bulbils, small clonal propagules produced on roots and in leaf axils, or rhizome buds (Eckert et al. 2000, 2003). Although populations in the native European range are thought to be largely triploid, in North America the diploid cytotype is most common (Kliber et al. 2005). In 2013, a programme to develop biological control agents for B. umbellatus was initiated in collaboration between North American and European scientists (www.cabi. org/projects/project/56422). To date, there have been limited ecological comparisons between introduced B. umbellatus cytotypes in North America, but there is strong evidence that cytotype identity (i.e. triploid or diploid) is related to disease resistance (Harms et al. 2020), and growth and reproductive success (Eckert et al. 2000; Lui et al. 2005). Other than a single study in the native range (Hroudová and Zákravský 1993), no one has investigated the role of ploidy in plant response to increased nutrient availability for B. umbellatus.

Glasshouse experiment

To test for genetic-based differences in available nutrientrelated growth and tissue chemistry plasticity in introduced *B. umbellatus* populations, we conducted a glasshouse common garden experiment. Plants were field-collected from 10 sites during 2016–17 (Fig. 1; **see Supporting Information—Table S1.1**), genotyped by Amplified Fragment Length Polymorphisms and grown in a common garden at the U.S. Army Engineer Research and Development Center (ERDC), Vicksburg, MS, USA. Plants from four triploid (Genotype 1; G1), and six diploid (four Genotype 4, one Genotype 3 and one Genotype 5; G4, G3, G5) populations were repeatedly vegetatively propagated. Our selection of source populations for this work was made with the goal to include populations from a large geographic area, in order to capture genetic and phenotypic variation present in the USA. Genetic data in Gaskin *et al.* (2021) suggest that both diploid and triploid plants in the North American invasion reproduce clonally and not from seed via outcrossing. Thus, genetic diversity in North America is limited to only six genotypes (G1–G6). There was no genetic diversity within our diploid or triploid source populations (i.e. they were clonal), and thus all plants of a designated genotype are genetically identical.

Plants were grown outdoors in commercially available topsoil amended with slow-release fertilizer (Osmocote®; 15-9-2; Scotts Miracle-Gro, Marysville, OH, USA) and propagated from cuttings at least twice over 2 years to reduce maternal effects (Roach and Wulff 1987). Water was municipal-delivered and maintained 10–20 cm above the sediment surface in cultures. Approximately 1.5 weeks before transplanting for the experiment, potted plants were harvested, rinsed of debris and then prepared for repotting. For each population, rhizome fragments (~4 cm long) with leaves removed were floated in aerated water until replanting.

Six nutrient solutions were prepared, containing varying amounts of N or P [see Supporting Information—Table S1.2]. Solution compositions were based on a standard Hoagland's recipe (200 mg L⁻¹ N; 40 mg L⁻¹ P) (Hoagland and Arnon 1950), but modified to create levels of N or P that varied along a logarithmic scale for the experiment [see Supporting Information—Supplement 1] (similar to Garrish *et al.* 2010). To test for B. *umbellatus* response to N, four solutions were used in which [P] was held constant at 40 mg L⁻¹ and [N] was modified (400, 200, 40, 4 mg L⁻¹). Likewise, three solutions were used in which [N] was constant at 40 mg L⁻¹ and [P] altered (40, 4, 0.4 mg L⁻¹). All other micro- and macronutrients were the same between solutions. The range of nutrients used in this study, although designed to measure plant response across a gradient, fall within the range used in previous work with B. *umbellatus* (Manolaki *et al.* 2020).

To test B. umbellatus phenotypic response to N or P, plants were grown hydroponically within a single glasshouse at

Figure 1. (A) Map of locations where B. umbellatus plants were collected for this work. Genotypes and ploidy are indicated by the symbols: White triangle = triploid G1, black circle = diploid G3, black square = diploid G4, black triangle = diploid G5. Also shown are (B) a close-up of a B. umbellatus inflorescence, and (C) a triploid B. umbellatus infestation in Oconto Falls, WI, USA.



the ERDC. At planting, plant propagules were weighed, then placed into net pots (12.7 mm diameter) filled with washed expanded clay rocks (8-16 mm diameter). Net pots were placed individually within white 4-L polyethylene food containers and charcoal-filtered tap water was added to each container. After 1 week, the water was drained and 1.5 L of nutrient solution was added to each container. One week after initial set-up, any plants that did not show signs of growth were replaced with new, pre-weighed propagules. We used six plant replicates per population for each treatment combination. Plants were grown for 8 weeks and nutrients were completely exchanged weekly. At harvest, plants were rinsed with reverse osmosis water, then separated into roots, shoots (leaves) and clonal reproductive (bulbils, rhizomes) tissues, placed into paper bags and dried at 60 °C. Only two plants produced inflorescences during the study, so sexual tissue biomass was not included in analyses. Tissues were weighed, then three randomly chosen replicates from each treatment combination were processed for tissue chemistry. Tissue chemistry analyses were performed at the Louisiana State University Agricultural Chemistry Laboratory, Baton Rouge, LA, USA. Nitrogen and carbon were determined by the modified Dumas method (CN 628 Dumas Analyzer; LECO, St. Joseph, MI, USA) and tissue phosphorus was determined by inductively coupled plasma (ICP) mass spectrometry (ARCOS; SPECTRO Analytical Instruments, Kleve, Germany) (Jones and Case 1990). For some samples, particularly those grown in the lowest N treatment (4 mg L⁻¹ N), there was insufficient material for all nutrient analyses. In those cases, we prioritized analyses of C and N over P. From the results of plant tissue nutrient analyses, C:N ratio, C:P ratio and N:P ratio were calculated.

Statistical approach

We first tested whether B. umbellatus biomass phenotypes (total biomass, reproductive biomass, shoot:root ratio) differed in response to N and P using separate mixed-effects models (general linear model [GLM]). Fixed effects for the models were cytotype (two levels: diploid, triploid), nutrient level (four levels of N, three levels of P; N and P treatments were tested separately), the cytotype * nutrient level interaction and initial fresh weight of propagules. We did not have replication at the level of genotypes so examined variation in cytotype only. Thus, we assigned population as a random effect in all models and initial (at planting) fresh weight of plants was included as a covariate. A significant interaction involving cytotype indicated that diploid and triploid B. umbellatus differed in their plasticity of the measured trait to nutrient enrichment. To quantify the direction and magnitude of plasticity differences, we calculated effects sizes, as described below.

Additionally, we tested whether tissue chemistry differed between plant parts, cytotypes and nutrient levels. We first generated standardized Z-scores (i.e. variables with mean = 0 and standard deviation = 1) for chemistry variables (%N, %C, %P, C:N ratio, N:P ratio, C:P ratio) to control for differences in variance and measurement units of each variable (Gotelli and Ellison 2004). Next, to reduce dimensionality in the data set and decrease the likelihood of committing type I errors, we conducted a principal component analysis (PCA) on the Z-scores. Principal components (PCs) with eigenvalues > 1 were retained for analysis. The resulting PCs represented independent linear combinations of tissue chemistry variables and accounted for 71 % (two PCs for nitrogen nutrient treatment) and 70 % (two PCs for phosphorus nutrient treatment) of total variance present in the original variables. Tissue chemistry PCs were tested separately with mixedeffects models, using cytotype, plant part, nutrient level and all two- and three-way interactions as fixed effects, population as a random effect and initial fresh weight of flowering rush propagules as a covariate. As above, significant interactions involving cytotype (i.e. cytotype * nutrient, cytotype * plant part, cytotype * nutrient * plant part) indicated significant differences in phenotypic plasticity between diploid and triploid flowering rush plants. Because there was limited biomass produced in triploid plants under the lowest (4 mg L-1) N treatment, we removed the low-N treatment from those analyses. For clarity, we limit discussion of tissue chemistry variables to those within each PC that were correlated (r > 0.5). Degrees of freedom for all mixed models were calculated following Kenward and Roger (1997). All statistical tests were performed with SAS ver. 9.4 (SAS Institute Inc., Cary, NC, USA, 2018).

Variation in phenotypic plasticity

We took two approaches to compare plasticity in measured biomass and leaf chemistry traits between cytotypes. In the first approach we simply determined whether the cytotype * nutrient level interaction was significant (P < 0.05) in models described in the previous section. A significant interaction would indicate that phenotypic responses of cytotypes differed in relation to nutrient level. Our second approach, consisted of using effect size to quantify differences in direction and magnitude of response between cytotypes. To compare phenotypic plasticity between the two cytotypes, we calculated Hedge's q (J-corrected Cohen's d) (Borenstein et al. 2011; Davidson et al. 2011) from population means for each trait and nutrient level. This type of approach is valuable because it allows for a direct comparison with multiple traits standardized in units of standard deviation (Cook-Patton and Agrawal 2011) and unlike some other commonly used indices of phenotypic plasticity (e.g. Valladares et al. 2006), its values are not restricted from 0 to 1, so it accounts for both direction and magnitude of plasticity. We calculated g for each population and nutrient combination and used populationlevel *q* to test for differences in plasticity between cytotypes. First, d was calculated as:

$$d = rac{Mean_{max} - Mean_{min}}{SD_{pooled}}$$

where Mean_{max} and Mean_{min} were the maximum and minimum mean values for each nutrient and population treatment. Pooled standard deviation was calculated as the square root of the mean of the two standard deviations (Cohen 1988):

$$SD_{pooled} = -\frac{\sqrt{SD_{max} + -SD_{min}}}{2}$$

We then corrected for small sample size by applying the *J* correction (Borenstein *et al.* 2011):

$$J = 1 - \frac{3}{(4df - 1)},$$
$$g = J * d,$$

with 10 degrees of freedom for biomass and four for tissue chemistry analyses. We present cytotype means and 95 % confidence intervals for *g*, generated from population-level calculations.

Results

Variation in biomass response to N and P between cytotypes

Despite our prediction that triploid plants would outperform diploid plants at increasing nutrient levels, diploid plants achieved greater total and reproductive biomass than triploid plants across nutrients (Figs 2 and 3; see Supporting Information—Table S2.1). Although both cytotypes performed similarly at low N, diploid plant biomass increased strongly with N (i.e. steep response curve), whereas triploid plant biomass increased, but at a lesser rate (resembling a 'master-of-some' scenario for the diploid B. umbellatus cytotype relationship between biomass variables and N; sensu Richards et al. 2006). Consequently, there were significant cytotype * N interactions for total biomass (P < 0.001), reproductive biomass (P < 0.001) and shoot:root ratio (P = 0.02) [see Supporting Information-Table S2.2], indicating significant differences in plasticity between the two cytotypes for these traits. At 400 mg L⁻¹ N, diploid plants produced nearly twice as much biomass overall and two and a half times more reproductive (bulbils) biomass, but 10 % less root biomass than triploid plants. Triploid plants produced 30 % more root biomass across N treatments, evident in the consistently lower calculated shoot:root ratios (Fig. 2E). Likewise, diploid plants consistently outperformed triploid plants in response to P additions (Fig. 3), resembling a 'jack-and-master' scenario for the relationship between biomass variables and P. We detected a significant cytotype * P interaction for total dry weight biomass (P < 0.001). Across levels of P, diploid plants produced 33 % more total biomass and 159 % more reproductive biomass than triploid plants (P = 0.003 and P < 0.001, respectively; see Supporting Information—Tables S2.1 and S2.2). However, shoot:root ratio of triploid plants was 68 % lower than diploid plants (Fig. 3E), a finding consistent with the higher root biomass allocation response to N.

The amount of variation in biomass response to N and P among populations depended on nutrient (Figs 2 and 3; see Supporting Information—Table S2.3). Triploid populationlevel variation was more than double diploid variation for reproductive biomass across N treatments (triploid coefficient of variation [CV] = 0.44, diploid CV = 0.18). However, diploid populations displayed 6 % (triploid CV = 0.16, diploid CV = 0.17) and 45 % (triploid CV = 0.08, diploid CV = 0.12) more variation in total biomass responses to N and P, respectively. Variation in shoot:root ratio was similar between cytotypes across levels of N (triploid CV = 0.37, diploid CV = 0.36), but across levels of



Figure 2. Reaction norms for *B. umbellatus* cytotypes (A, C, E) and populations (B, D, F). Reported are the means ± SE for growth and reproductive responses to four levels of nitrogen. Within cytotypes, we used four triploid populations of a single genotype and six diploid populations of three different genotypes (four G4, one G3 and one G5 population). DW = dry weight.



Figure 3. Reaction norms for *B. umbellatus* cytotypes (A, C, E) and populations (B, D, F). Reported are the means \pm SE for growth and reproductive responses to three levels of phosphorus. Within cytotypes, we used four triploid populations of a single genotype and six diploid populations of three different genotypes (four G4, one G3 and one G5 population). DW = dry weight.

P diploid populations were nearly twice as variable (triploid CV = 0.16, diploid CV = 0.30). Increased variation among diploid populations may be due to the increased genetic diversity of diploid cytotypes in the introduced range. Although there is a sole triploid genotype currently known in North America, there are a number of diploid genotypes. The G4 B. umbellatus genotype is the most common diploid B. umbellatus genotype in North America (Gaskin et al. 2021) and all G4 populations (n = 4) had similar responses to nutrient enrichment (Figs 2 and 3). The lone G5 population (Forest Lake, MN, USA) had a similar response curve to the other diploid populations. An interesting finding is that the Springbrook Pond, IL population (G3) had traits of both diploid and triploid populations. For instance, G3 plants were more similar to triploid than other diploid plants in shoot:root ratio, but not other measured responses to high levels of P or N (Figs 2 and 3). G3 plants responded to increased N and P with higher reproductive output and total biomass production than triploid plants and were similar in that respect to other diploid populations. Thus, G3 plants had an intermediate response to increased nutrients when compared with triploid and other diploid populations. However, we acknowledge that because of our sample size for diploid genotypes (n = 1 population for each) G3 and G5), we were unable to fully investigate variation at the genotype level.

Variation in tissue chemistry in response to N and P

Tissue chemistry varied by cytotype, level and type of nutrient availability, and in allocation to plant parts (Fig. 4; see Supporting Information—Tables S2.1 and S2.5). We generated two PCs (PC1_N, PC2,) that explained 39 % and 32 %, respectively, of variation in tissue chemistry (combined 71 %) in response to different levels of N (see supporting information Table S2.4). PC1_N was negatively correlated to tissue phosphorus (r = -0.61) and positively correlated with C:P ratio (r = 0.89) and N:P ratio (r = 0.90). PC2_N was positively correlated with tissue nitrogen (r = 0.86) and negatively correlated with C:N ratio (r = -0.83). Similarly, two PCs (PC1_p, PC2_p) that explained 40 % and 29 % of variation in tissue chemistry (combined 69 %) were generated from phosphorus-response variables. $PC1_{p}$ was positively correlated with N (r = 0.60), N:P ratio (r = 0.85) and C:P ratio (r = 0.87), and negatively correlated with C:N ratio (r = -0.51). PC2_p was negatively correlated with N (r = -0.77) and positively with C:N ratio (r = 0.79). Overall, the pattern for both sets of PCs was as follows: $\mathrm{PC1}_{_{\mathrm{N}}}$ was most related to tissue phosphorus, $PC2_{N}$ and $PC2_{P}$ were most related to tissue nitrogen and $\text{PC1}_{\scriptscriptstyle \rm D}$ was related to both tissue nitrogen and phosphorus.



Figure 4. Mean ± SE tissue chemistry PCs for flowering rush diploid or triploid plant parts, grown in increasing levels of nitrogen (A, B) or phosphorus (C, D). The individual variable loadings for the PCs are shown in adjacent panels.

We found differences in tissue chemistry PCs between cytotypes, although many differences depended on nutrient level and plant part (Fig. 4; **see Supporting Information—Table S2.5**). Overall, tissue nitrogen ($PC2_N$, $PC1_p$) was higher in triploid plants than diploid plants across nitrogen ($PC2_N$; P = 0.03) or phosphorus ($PC1_p$; P = <0.001) treatments, and C:N ratio was consistently lower in diploid versus triploid plants (Fig. 4). Not surprisingly, N:P ratio ($PC1_N$, $PC1_p$) generally increased with nitrogen (P = 0.003) and decreased with phosphorus (P < 0.001) nutrients for both cytotypes, although based on means of untransformed data, N:P ratios in triploid plants declined 89 % between 200 and 400 mg L⁻¹ N, whereas N:P in diploid plants increased 236 % over the same nutrient interval.

Although root N and P concentrations in both cytotypes increased with N, triploid plants typically had higher root concentrations overall (Fig. 4B). Triploid plants had 38 % higher P (0.64 % dry weight P) and 13 % higher N (3.67 % dry weight N) but 32 % lower C:N ratio overall in roots. In response to variation in P, roots in triploid plants had higher concentrations of N (46 % higher), P (52 % higher) and C (4 % higher), but a 34 % lower C:N ratio than diploid plants. In shoots, diploid plants had 5 % higher C and 24 % higher C:N ratio, but lower N and P (both 11 % lower). In response to P, reproductive tissue chemistry was similar between cytotypes for C and P but triploid plants had 18 % higher N and 15 % lower C:N than diploid plants.

Phenotypic plasticity in response to N and P

We detected differences in plasticity to nutrients, as a biomass response, between the triploid and diploid cytotypes (Fig. 5). Contrary to our prediction, diploid plants displayed two times the plasticity for total biomass (diploid g = 5.35, triploid g = 2.77) and more than double the plasticity in reproductive biomass (diploid g = 4.87; triploid g = 1.84) in response to N. Plasticity in shoot:root ratio was nearly three times higher for diploid plants (g = 1.92) in response to N than triploid plants (g = 0.62). In response to



Figure 5. Mean \pm 95 % CI for phenotypic plasticity (Hedge's *g*) of introduced *B. umbellatus* cytotypes in response to N or P enrichment. TTL = total biomass, RE = reproductive biomass, S:R = the ratio of shoot (leaf) to root biomass.

P, diploid plant plasticity was double that of triploid plants for total biomass (diploid g = 1.85, triploid g = 0.97), reproductive biomass (diploid g = 2.61, triploid g = 1.16) and shoot:root ratio (diploid g = 1.51, triploid g = 0.70).

Plasticity of tissue chemistry varied between plant parts but we observed no consistent cytotype differences (Fig. 6). Plasticity of chemistry N and C:N ratio PCs was highest for shoot (PC2_N; diploid g = 2.86, triploid g = 1.99) and root (PC2_N: diploid g = 2.39, triploid g = 1.94) tissues in response to N, and shoots (PC1_P: diploid g = 1.08, triploid g = 1.14; PC2_P: diploid g = 1.12, triploid g = 1.30) in response to P. Plasticity of the P PC (PC1_N) was highest



Figure 6. Plasticity (Hedge's g) in tissue chemistry PCs of introduced B. umbellatus cytotypes in response to P or N enrichment. Points are mean values and error bars are 95 % confidence intervals.

(diploid g = 2.16, triploid g = 1.85) in shoots in response to N. For both cytotypes, plasticity in shoot PC1_N (P, C:P ratio, N:P ratio), shoot and root PC2_N (N, C:N ratio), shoot and reproductive PC2_P (N, C:N ratio) and reproductive PC1_P (N, C:P ratio, N:P ratio) were greater than zero (Fig. 6).

Discussion

Our results demonstrate that different cytotypes present in B. umbellatus populations within North America display significant phenotypic variation in traits, and support a growing list of studies that have shown genetic-based variation in character traits among invader populations that correlate with invasion success (Eckert et al. 2000; Lavergne and Molofsky 2007). For instance, the weed Phyla canescens has been introduced into several areas outside its native range, but the traits thought to be most important for its success vary between invaded areas, with high seed output in Australia but vigorous clonal/ vegetative growth in France (Xu et al. 2010). In the current study, and contrary to our prediction, increased ploidy in B. umbellatus was not associated with greater plasticity to increased nutrients and, although there was considerable variation among diploid populations, diploid populations consistently outperformed triploid populations across two nutrient gradients. Although less plastic for measured traits in response to variation in nutrient levels, triploid B. umbellatus has invaded aquatic systems across the USA and Canada, creating expansive monocultures and causing negative economic (e.g. water delivery), recreational (Boutwell 1990), ecological (Parkinson et al. 2010) and human health (e.g. cercarial dermatitis; Parkinson et al. 2010) impacts. In triploid plants, all traits except below-ground biomass plateaued in our moderately low N treatment (40 mg L-1). At higher N concentrations, the only significant response observed in triploid plants occurred as an increase in root biomass and related lower shoot:root ratio. This is surprising because allocation of resources to underground biomass should typically occur when nutrients are scarce (Poorter et al. 2012). However, below-ground biomass and competitive ability can be positively correlated (Aerts et al. 1991; Cahill et al. 2000), so an increase in below-ground biomass may be adaptive and thus important in explaining invasion success and negative impacts observed in recipient communities. Furthermore, as below-ground competition is expected to be highest in nutrient-limited environments (Tilman 1989; Cahill 1999), triploid B. umbellatus plants with low shoot:root ratios may be better adapted than diploid plants to compete for soil resources in low nutrient environments. Studies that address the question of whether polyploids of invasive species perform better in low-resource environments than their diploid relatives are relatively rare, but polyploidy in general may lead to increased performance in extreme conditions, such as at high altitudes or xeric environments (del Pozo and Ramirez-Parra 2015), and increase tolerance to nutrient or water stress (Deng et al. 2012; Allario et al. 2013; but see discussion of the Large Genome Constraint Hypothesis below for a contrasting viewpoint).

Increased biomass, reproductive, or chemical plasticity in response to nutrient gradients has been shown in other systems to be indicative of invasiveness (Burns 2004), or resistance to invasion (Peacor et al. 2006), and is typically associated with higher ploidy (Levin 1983; Leitch and Leitch 2008; te Beest et al. 2012) and smaller genome size (e.g. Knight et al. 2005; Hessen et al. 2010). Although the Large Genome Constraint Hypothesis posits that the metabolic costs of replicating excess DNA may make polyploidy maladaptive for an organism, especially in extreme environments (Knight et al. 2005), costs of polyploidy are likely outweighed by the benefits, which may include positive effects of gene redundancy on masking deleterious alleles and adaptive functional divergence of the replicate genes (te Beest et al. 2012). In the current study, diploid plants displayed higher plasticity of measured traits to a nutrient gradient. Greater biomass and reproductive plasticity to nutrients may partially explain why their distributional range in North America is broader than the range for triploid B. umbellatus (widely distributed across north-eastern and Midwestern states as compared to the Pacific Northwest). With one exception, diploid and triploid plants are not known to co-occur in the USA, but this study raises the question of potential outcomes of co-invasion by cytotypes in the same area and whether diploid plants will outperform triploid plants in field locations, ultimately displacing triploid plants and becoming the dominant cytotype. Investigations into competitive interactions between genotypes of *B. umbellatus*, particularly along resource gradients, may lead to further predictions on which habitats are at risk for invasion. Though it has not been reported, a possible explanation for the success of both cytotypes in the USA, and that they do not frequently co-occur, is that they occupy habitats with differing nutrient availability.

The implications of variation in growth and biomass allocation plasticity between genotypes or cytotypes extend beyond nutrient uptake and possible competitive interactions with other plant species to potentially impact trophic interactions. Performance of herbivores is influenced by host plant abundance (biomass available for food), nutritional status (N, C:N) (Scriber and Slansky 1981; Awmack and Leather 2002; Wilson et al. 2007; Harms and Cronin 2019), allocation patterns of nutrients within plant parts (Eatough Jones et al. 2008; Hunter 2016), defensive chemistry (Throop and Lerdau 2004; Nybakken et al. 2018) and interactions between plants and higher tropic levels (i.e. mediated by plant volatiles; Turlings and Erb 2018), all of which are affected by plant phenotypic response to nutrients. Greater growth or tissue chemistry plasticity in response to herbivory may increase plant performance through tolerance (i.e. reduced effects of herbivory on plant fitness) or resistance (e.g. reduced performance of the herbivore through interaction with plant chemical defences) (Yoshizuka and Roach 2011). Variation in induced resistance (the defensive reaction by a plant to damage from an herbivore or pathogen) may be particularly important for invasive plants because the magnitude and type of induction can depend on the identity, type or origin of herbivore encountered (Liu et al. 2018; Zhang et al. 2018). Herbivores or pathogens introduced for management (i.e. biological control) will be particularly susceptible to variation in plant plasticity because biological control agents are restricted to a single host species.

Nutrient availability via the bottom-up cascade from environment to host plant to herbivore can ultimately determine success of biological control (Room and Thomas 1985; Room et al. 1989; Wheeler and Center 1996; Center and Dray 2010; Hunter 2016; Uyi et al. 2016). For example, fertilizer application was needed to increase tissue N in giant salvinia (Salvinia molesta) to levels that supported egg production and larval development of the biological control agent, Cyrtobagous salviniae (Room and Thomas 1985). Similarly, increasing N availability in nutrient solution led to higher tissue N and lower tissue C:N ratio in A. philoxeroides, accelerated larval development and larger adults of the biological control agent Agasicles hygrophila (Harms and Cronin 2019). Thus, an understanding of relationships between environment and plant nutrients is important when testing efficacy of potential biological control agents or evaluating establishment and control failures at field sites (Room and Thomas 1985; Wheeler and Center 1997). In general, herbivore performance is positively correlated with tissue N and negatively with tissue C:N ratio (Awmack and Leather 2002), a pattern which has also been found for a generalist herbivore on B. umbellatus (Harms and Walter 2021). In the current study, N and C:N ratio varied between cytotypes in several instances; C:N ratio was lower in triploid plants overall in response to N and P enrichment. Thus, triploid B. umbellatus plants with lower C:N ratio may be more susceptible to herbivory than diploid plants and may therefore be more affected by biological control,

if introduced. In support of this argument, Harms *et al.* (2020) found that triploid *B. umbellatus* plants were more susceptible than diploid plants to infection by generalist pathogens, at least under controlled experimental conditions. Additionally, Harms and Walter (2021) found decreased N, and increased C:N ratio, of diploid versus triploid *B. umbellatus* plants likely contributed to reduced herbivore performance in laboratory assays. Mindful of potential differences in plant growth and phenotype between greenhouse and field conditions (e.g. Poorter *et al.* 2016), we suggest that additional studies be undertaken to better understand nutrient conditions in field infestations of diploid and triploid *B. umbellatus*.

For genetically diverse invaders, genetic identity and associated phenotypic responses to environmental conditions can be important for understanding ecological impacts resulting from establishment and spread (Lee 2002; terHorst and Lau 2015) or for making management decisions to curb future impacts (Gaskin et al. 2011). In the wetland invader, B. umbellatus, we found differences in phenotypic plasticity related to enrichment with N or P. In contrast to a previous report (Hroudová and Zákravský 1993), we found that triploid B. umbellatus plants underperformed relative to diploid plants in all biomass categories except roots and were significantly less plastic in response to available nutrients. Whether disproportionate allocation of biomass to roots is advantageous for triploid plants and thus contributing to invasion success in the USA is unknown. However, the clear differences between cytotypes in biomass allocation and chemistry suggest that cytotypes can occupy different habitats and may require unique management approaches.

Supporting Information

The following additional information is available in the online version of this article—

Table S1.1. Populations used in the current study to detect differences in *Butomus umbellatus* cytotype biomass and chemistry responses to gradients of N or P.

Table S1.2. Nutrient solutions used in the current experiment. Approximate concentrations (mg L^{-1}) of macronutrients given. Micronutrients followed the standard Hoagland's nutrient recipe (Hoagland and Arnon 1950).

Table S2.1. Least squared means and standard errors for diploid and triploid flowering rush plant biomass and tissue chemistry variables in relation to nutrient solution (both biomass and tissue chemistry) and plant part (tissue chemistry only). Means were calculated using mixed-effects models (GLM). Fixed effects for the biomass (total dry weight, shoot:root ratio, reproductive dry weight) models were cytotype (two levels: diploid, triploid), nutrient level (four levels of N, three levels of P; N and P treatments were tested separately), the cytotype * nutrient level interaction and initial fresh weight of propagules. We did not have replication at the level of genotypes so examined variation in cytotype only. Thus, we assigned population as a random effect in all models and initial (at planting) fresh weight of plants was included as a covariate. For tissue chemistry (%C, %N, %P, C:N ratio, C:P ratio, N:P ratio), we used similar mixed models as above. Fixed effects were cytotype (two levels: diploid, triploid), nutrient level (four levels of N, three levels of P; N and P treatments were tested separately), plant part (shoot, root, reproductive parts) and all two- and three-way interactions. As above, population was a random effect in all models and initial (at planting) fresh weight of plants was included as a covariate.

Table S2.2. Test statistics from mixed-effects models for growth and reproductive biomass responses of diploid and triploid *Butomus umbellatus* plants to N and P. Models are described above.

Table S2.3. Coefficients of variation (CV), calculated for diploid and triploid cytotype biomass variables, using population means.

Table S2.4. Individual variable loadings for each Butomus umbellatus tissue chemistry principal component (PC). PCs with eigenvalues > 1 were retained for use in mixed models.

Table S2.5. Mixed model test statistics for tissue chemistry principal components (PCs) of diploid and triploid Butomus umbellatus plant response to N and P. Tissue chemistry PCs were tested separately with mixed-effects models, using cytotype, plant part, nutrient level and all two- and three-way interactions as fixed effects, B. umbellatus population as a random effect and initial fresh weight of flowering rush propagules as a covariate. Degrees of freedom for all mixed models were calculated following Kenward and Roger (1997).

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Conflict of Interest

None declared.

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Contributions by the Authors

N.E.H. performed the research, secured the funding, collected data and conducted data analysis. N.E.H., J.T.C. and J.F.G. interpreted the data and wrote the manuscript.

Data Availability

The data that support the findings of this study are available from the corresponding author upon reasonable request or at https://doi.org/10.6084/m9.figshare.14935725.

Literature Cited

- Aerts R, Boot RG, van der Aart PJ. 1991. The relation between above- and belowground biomass allocation patterns and competitive ability. Oecologia 87:551–559.
- Agrawal AA, Conner JK, Johnson MT, Wallsgrove R. 2002. Ecological genetics of an induced plant defense against herbivores: additive genetic variance and costs of phenotypic plasticity. Evolution 56:2206–2213.
- Allario T, Brumos J, Colmenero-Flores JM, Iglesias DJ, Pina JA, Navarro L, Talon M, Ollitrault P, Morillon R. 2013. Tetraploid Rangpur lime rootstock increases drought tolerance via enhanced constitutive root abscisic acid production. Plant, Cell & Environment 36:856–868.
- Anderson LC, Zeis CD, Alam SF. 1974. Phytogeography and possible origins of Butomus in North America. Bulletin of the Torrey Botanical Club 101:292–296.
- Ashton IW, Lerdau MT. 2008. Tolerance to herbivory, and not resistance, may explain differential success of invasive, naturalized, and native North American temperate vines. *Diversity and Distributions* 14:169–178.

- Awmack CS, Leather SR. 2002. Host plant quality and fecundity in herbivorous insects. *Annual Review of Entomology* **47**:817–844.
- Bhattarai GP, Meyerson LA, Anderson J, Cummings D, Allen WJ, Cronin JT. 2017. Biogeography of a plant invasion: genetic variation and plasticity in latitudinal clines for traits related to herbivory. *Ecological Monographs* 87:57–75.
- Blumenthal DM. 2006. Interactions between resource availability and enemy release in plant invasion. Ecology Letters **9**:887–895.
- Borenstein M, Hedges LV, Higgins JP, Rothstein HR. 2011. Introduction to metaanalysis. Chichester, West Sussex, United Kingdom: John Wiley & Sons.
- Bossdorf O, Auge H, Lafuma L, Rogers WE, Siemann E, Prati D. 2005. Phenotypic and genetic differentiation between native and introduced plant populations. *Oecologia* 144:1–11.
- Boutwell JE. 1990. Flowering-rush: a plant worth watching. Aquatics 9:8-11.
- Burns JH. 2004. A comparison of invasive and non-invasive dayflowers (Commelinaceae) across experimental nutrient and water gradients. Diversity and Distributions 10:387–397.
- Cahill JF Jr. 1999. Fertilization effects on interactions between above- and belowground competition in an old field. *Ecology* **80**:466–480.
- Cahill J, James F, Casper BB. 2000. Investigating the relationship between neighbor root biomass and belowground competition: field evidence for symmetric competition belowground. *Oikos* **90**:311–320.
- Center TD, Dray FA. 2010. Bottom-up control of water hyacinth weevil populations: do the plants regulate the insects? *Journal of Applied* Ecology **47**:329–337.
- Chevin LM, Lande R. 2011. Adaptation to marginal habitats by evolution of increased phenotypic plasticity. *Journal of Evolutionary Biology* 24:1462–1476.
- Cohen J. 1988. Statistical power analysis for the behavioral sciences. New York, NY: Lawrence Erlbaum Associates, Publishers.
- Cook-Patton SC, Agrawal AA. 2011. Relatedness predicts phenotypic plasticity in plants better than weediness. *Evolutionary Ecology Research* **13**:527–542.
- Davidson AM, Jennions M, Nicotra AB. 2011. Do invasive species show higher phenotypic plasticity than native species and, if so, is it adaptive? A meta-analysis. *Ecology Letters* **14**:419–431.
- del Pozo JC, Ramirez-Parra E. 2015. Whole genome duplications in plants: an overview from Arabidopsis. Journal of Experimental Botany 66:6991–7003.
- Deng B, Du W, Liu C, Sun W, Tian S, Dong H. 2012. Antioxidant response to drought, cold and nutrient stress in two ploidy levels of tobacco plants: low resource requirement confers polytolerance in polyploids? *Plant Growth Regulation* **66**:37–47.
- DeWalt SJ, Denslow JS, Hamrick JL. 2004. Biomass allocation, growth, and photosynthesis of genotypes from native and introduced ranges of the tropical shrub Clidemia hirta. Oecologia **138**:521–531.
- Eatough Jones M, Paine TD, Fenn ME. 2008. The effect of nitrogen additions on oak foliage and herbivore communities at sites with high and low atmospheric pollution. *Environmental Pollution* **151**:434–442.
- Eckert CG, Lui K, Bronson K, Corradini P, Bruneau A. 2003. Population genetic consequences of extreme variation in sexual and clonal reproduction in an aquatic plant. *Molecular Ecology* **12**:331–344.
- Eckert CG, Massonnet B, Thomas JJ. 2000. Variation in sexual and clonal reproduction among introduced populations of flowering rush, *Butomus umbellatus* (Butomaceae). *Canadian Journal of Botany* **78**:437–446.
- Eller F, Brix H. 2012. Different genotypes of Phragmites australis show distinct phenotypic plasticity in response to nutrient availability and temperature. Aquatic Botany **103**:89–97.
- Fan S, Liu C, Yu D, Xie D. 2013. Differences in leaf nitrogen content, photosynthesis, and resource-use efficiency between Eichhornia crassipes and a native plant Monochoria vaginalis in response to altered sediment nutrient levels. Hydrobiologia 711:129–137.
- Flanagan NE, Richardson CJ, Ho M. 2015. Connecting differential responses of native and invasive riparian plants to climate change and environmental alteration. *Ecological Applications* **25**:753–767.
- Funk JL. 2013. The physiology of invasive plants in low-resource environments. Conservation Physiology 1:1–17.
- Garrish V, Cernusak LA, Winter K, Turner BL. 2010. Nitrogen to phosphorus ratio of plant biomass versus soil solution in a tropical pioneer tree, Ficus insipida. Journal of Experimental Botany 61:3735–3748.

- Gaskin JF, Andreas J, Grewell BJ, Haefliger P, Harms NE. 2021. Diversity and origins of Butomus umbellatus (flowering rush) invasion in North America. Aquatic Botany 173:103400.
- Gaskin JF, Bon M-C, Cock MJ, Cristofaro M, De Biase A, De Clerck-Floate R, Ellison CA, Hinz HL, Hufbauer RA, Julien MH. 2011. Applying molecularbased approaches to classical biological control of weeds. *Biological* Control 58:1–21.
- Gaskin JF, Schwarzländer M, Kinter CL, Smith JF, Novak SJ. 2013. Propagule pressure, genetic structure, and geographic origins of *Chondrilla juncea* (Asteraceae): an apomictic invader on three continents. *American Journal of Botany* **100**:1871–1882.
- Gioria M, Osborne BA. 2014. Resource competition in plant invasions: emerging patterns and research needs. Frontiers in Plant Science 5:501.
- Godoy O, Valladares F, Castro-Díez P. 2011. Multispecies comparison reveals that invasive and native plants differ in their traits but not in their plasticity. *Functional Ecology* **25**:1248–1259.
- Golani D, Azzurro E, Corsini-Foka M, Falautano M, Andaloro F, Bernardi G. 2007. Genetic bottlenecks and successful biological invasions: the case of a recent Lessepsian migrant. Biology Letters 3:541–545.
- Gotelli N, Ellison GN. 2004. A primer of ecological statistics. Sunderland, MA: Sinauer Associates, Inc.
- Grewell BJ, Skaer Thomason MJ, Futrell CJ, Iannucci M, Drenovsky RE. 2016. Trait responses of invasive aquatic macrophyte congeners: colonizing diploid outperforms polyploid. AoB PLANTS 8:plw014; doi:10.1093/ aobpla/plw014.
- Griffith TM, Sultan SE. 2005. Shade tolerance plasticity in response to neutral vs green shade cues in *Polygonum* species of contrasting ecological breadth. *The New Phytologist* **166**:141–147.
- Gross KL, Mittelbach GG, Reynolds HL. 2005. Grassland invasibility and diversity: responses to nutrients, seed input and disturbance. *Ecology* **86**:476–486.
- Hahn MA, van Kleunen M, Müller-Schärer H. 2012. Increased phenotypic plasticity to climate may have boosted the invasion success of polyploid *Centaurea stoebe*. PLoS One 7:e50284.
- Hao GY, Lucero ME, Sanderson SC, Zacharias EH, Holbrook NM. 2013. Polyploidy enhances the occupation of heterogeneous environments through hydraulic related trade-offs in Atriplex canescens (Chenopodiaceae). The New Phytologist 197:970–978.
- Harms NE, Cronin JT. 2019. Variability in weed biological control: effects of foliar nitrogen on larval development and dispersal of the alligatorweed flea beetle, *Agasicles hygrophila*. Biological Control **135C**:16–22.
- Harms NE, Shearer JF, Cronin JT, Gaskin JF. 2020. Geographic and genetic variation in susceptibility of Butomus umbellatus to foliar fungal pathogens. Biological Invasions 22:535–548.
- Harms NE, Walter DJ. 2021. Influence of Butomus umbellatus L. lineage and age on leaf chemistry and performance of a generalist caterpillar. Aquatic Botany **172**:103391.
- Hastwell GT, Panetta FD. 2005. Can differential responses to nutrients explain the success of environmental weeds? *Journal of Vegetation Science* **16**:77–84.
- Hessen DO, Jeyasingh PD, Neiman M, Weider LJ. 2010. Genome streamlining and the elemental costs of growth. Trends in Ecology & Evolution **25**:75–80.
- Higgins SI, Richardson DM. 2014. Invasive plants have broader physiological niches. Proceedings of the National Academy of Sciences of the United States of America 111:10610–10614.
- Hoagland DR, Arnon DI. 1950. The water-culture method for growing plants without soil. Circular. California Agricultural Experiment Station 347:32pp.
- Hroudová Z, Zákravský P. 1993. Ecology of two cytotypes of Butomus umbellatus II. Reproduction, growth and biomass production. Folia Geobotanica 28:413–424.
- Huenneke LF, Hamburg SP, Koide R, Mooney HA, Vitousek PM. 1990. Effects of soil resources on plant invasion and community structure in Californian serpentine grassland. Ecology 71:478–491.
- Hunter MD. 2016. The phytochemical landscape: linking trophic interactions and nutrient dynamics. Princeton, New Jersey: Princeton University Press.
- Jones JB Jr, Case VW. 1990. Sampling, handling and analyzing plant tissue samples. In: Westerman RL, ed. Soil testing and plant analysis. Madison, WI: Soil Science Society of America, 389–427.
- Kay SH. 1992. Response of two alligatorweed biotypes to quinclorac. Journal of Aquatic Plant Management **30**:35–40.

- Kay SH, Haller WT. 1982. Evidence for the existence of distinct alligator weed biotypes. Journal of Aquatic Plant Management 20:37–41.
- Kenward MG, Roger JH. 1997. Small sample inference for fixed effects from restricted maximum likelihood. *Biometrics* **53**:983–997.
- Kettenring KM, McCormick MK, Baron HM, Whigham DF. 2011. Mechanisms of Phragmites australis invasion: feedbacks among genetic diversity, nutrients, and sexual reproduction. Journal of Applied Ecology 48:1305–1313.
- Kliber A, Eckert CG. 2005. Interaction between founder effect and selection during biological invasion in an aquatic plant. *Evolution* **59**:1900–1913.
- Knight CA, Molinari NA, Petrov DA. 2005. The large genome constraint hypothesis: evolution, ecology and phenotype. Annals of Botany 95:177–190.
- Kolb A, Alpert P. 2003. Effects of nitrogen and salinity on growth and competition between a native grass and an invasive congener. Biological Invasions 5:229–238.
- Lavergne S, Molofsky J. 2007. Increased genetic variation and evolutionary potential drive the success of an invasive grass. Proceedings of the National Academy of Sciences of the United States of America 104:3883–3888.
- Lee CE. 2002. Evolutionary genetics of invasive species. Trends in Ecology & Evolution 17:386–391.
- Leishman MR, Thomson VP. 2005. Experimental evidence for the effects of additional water, nutrients and physical disturbance on invasive plants in low fertility Hawkesbury Sandstone soils, Sydney, Australia. *Journal of Ecology* **93**:38–49.
- Leitch AR, Leitch IJ. 2008. Genomic plasticity and the diversity of polyploid plants. Science **320**:481–483.
- Levin DA. 1983. Polyploidy and novelty in flowering plants. The American Naturalist **122**:1–25.
- Liu M, Zhou F, Pan X, Zhang Z, Traw MB, Li B. 2018. Specificity of herbivoreinduced responses in an invasive species, Alternanthera philoxeroides (alligator weed). Ecology and Evolution 8:59–70.
- Lowe PN, Lauenroth WK, Burke IC. 2002. Effects of nitrogen availability on the growth of native grasses exotic weeds. *Journal of Range Management* **55**:94–98.
- Lui K, Thompson FL, Eckert CG. 2005. Causes and consequences of extreme variation in reproductive strategy and vegetative growth among invasive populations of a clonal aquatic plant, Butomus umbellatus L. (Butomaceae). Biological Invasions 7:427–444.
- Madsen JD, Wersal RM, Marko MD. 2016. Distribution and biomass allocation in relation to depth of flowering rush (Butomus umbellatus) in the Detroit Lakes, Minnesota. Invasive Plant Science and Management 9:161–170.
- Manolaki P, Tooulakou G, Byberg CU, Eller F, Sorrell BK, Klapa MI, Riis T. 2020. Probing the response of the amphibious plant *Butomus umbellatus* to nutrient enrichment and shading by integrating eco-physiological with metabolomic analyses. *Frontiers in Plant Science* **11**:1–14.
- Meyerson LA, Cronin JT, Bhattarai GP, Brix H, Lambertini C, Lučanová M, Rinehart S, Suda J, Pyšek P. 2016. Do ploidy level and nuclear genome size and latitude of origin modify the expression of *Phragmites australis* traits and interactions with herbivores? *Biological Invasions* **18**:2531–2549.
- Meyerson LA, Pyšek P, Lučanová M, Wigginton S, Tran C-T, Cronin JT. 2020. Plant genome size influences stress tolerance of invasive and native plants via plasticity. *Ecosphere* 11:e03145.
- Nybakken L, Lie MH, Julkunen-Tiitto R, Asplund J, Ohlson M. 2018. Fertilization changes chemical defense in needles of mature Norway spruce (Picea abies). Frontiers in Plant Science **9**:770.
- Pan XY, Jia X, Chen JK, Li B. 2012. For or against: the importance of variation in growth rate for testing the EICA hypothesis. *Biological Invasions* 14:1–8.
- Pan JJ, Price JS. 2001. Fitness and evolution in clonal plants: the impact of clonal growth. Evolutionary Ecology 15:583–600.
- Parkinson H, Mangold J, Dupuis V, Rice P. 2010. Biology, ecology and management of flowering rush (Butomus umbellatus). Bozeman, MT: Montana State University.
- Peacor SD, Allesina S, Riolo RL, Pascual M. 2006. Phenotypic plasticity opposes species invasions by altering fitness surface. PLoS Biology 4:e372.

- Peperkorn R, Werner C, Beyschlag W. 2005. Phenotypic plasticity of an invasive acacia versus two native Mediterranean species. Functional Plant Biology 32:933–944.
- Poorter H, Fiorani F, Pieruschka R, Wojciechowski T, van der Putten WH, Kleyer M, Schurr U, Postma J. 2016. Pampered inside, pestered outside? Differences and similarities between plants growing in controlled conditions and in the field. The New Phytologist 212:838–855.
- Poorter H, Niklas KJ, Reich PB, Oleksyn J, Poot P, Mommer L. 2012. Biomass allocation to leaves, stems and roots: meta-analyses of interspecific variation and environmental control. The New Phytologist 193:30–50.
- Richards CL, Bossdorf O, Muth NZ, Gurevitch J, Pigliucci M. 2006. Jack of all trades, master of some? On the role of phenotypic plasticity in plant invasions. Ecology Letters 9:981–993.
- Richards CL, Pennings SC, Donovan LA. 2005. Habitat range and phenotypic variation in salt marsh plants. Plant Ecology **176**:263–273.
- Richardson DM, Pyšek P. 2006. Plant invasions: merging the concepts of species invasiveness and community invasibility. Progress in Physical Geography: Earth and Environment 30:409–431.
- Riis T, Lambertini C, Olesen B, Clayton JS, Brix H, Sorrell BK. 2010. Invasion strategies in clonal aquatic plants: are phenotypic differences caused by phenotypic plasticity or local adaptation? Annals of Botany 106:813–822.
- Roach DA, Wulff RD. 1987. Maternal effects in plants. Annual Review of Ecology and Systematics 18:209–235.
- Room P, Julien M, Forno I. 1989. Vigorous plants suffer most from herbivores: latitude, nitrogen and biological control of the weed Salvinia molesta. Oikos 54:92–100.
- Room PM, Thomas PA. 1985. Nitrogen and establishment of a beetle for biological control of the floating weed salvinia in Papua New Guinea. *Journal of Applied Ecology* 22:139–156.
- Sax DF, Stachowicz JJ, Gaines SD. 2005. Species invasions: insights into ecology, evolution and biogeography: Sunderland, Massachusetts: Sinauer Associates Incorporated.
- Schierenbeck KA, Mack RN, Sharitz RR. 1994. Effects of herbivory on growth and biomass allocation in native and introduced species of Lonicera. Ecology 75:1661–1672.
- Scriber J, Slansky F Jr. 1981. The nutritional ecology of immature insects. Annual Review of Entomology 26:183–211.
- Seabloom EW, Borer ET, Buckley YM, Cleland EE, Davies KF, Firn J, Harpole WS, Hautier Y, Lind EM, MacDougall AS, Orrock JL, Prober SM, Adler PB, Anderson TM, Bakker JD, Biederman LA, Blumenthal DM, Brown CS, Brudvig LA, Cadotte M, Chu C, Cottingham KL, Crawley MJ, Damschen EI, Dantonio CM, DeCrappeo NM, Du G, Fay PA, Frater P, Gruner DS, Hagenah N, Hector A, Hillebrand H, Hofmockel KS, Humphries HC, Jin VL, Kay A, Kirkman KP, Klein JA, Knops JM, La Pierre KJ, Ladwig L, Lambrinos JG, Li Q, Li W, Marushia R, McCulley RL, Melbourne BA, Mitchell CE, Moore JL, Morgan J, Mortensen B, O'Halloran LR, Pyke DA, Risch AC, Sankaran M, Schuetz M, Simonsen A, Smith MD, Stevens CJ, Sullivan L, Wolkovich E, Wragg PD, Wright J, Yang L. 2015. Plant species' origin predicts dominance and response to nutrient enrichment and herbivores in global grasslands. *Nature Communications* 6:7710.
- Siebenkäs A, Schumacher J, Roscher C. 2015. Phenotypic plasticity to light and nutrient availability alters functional trait ranking across eight perennial grassland species. AoB PLANTS 7:plv029; doi:10.1093/aobpla/plv029.
- Simoes M, Baruch Z. 1991. Responses to simulated herbivory and water stress in two tropical C4 grasses. Oecologia 88:173–180.
- Steinger T, Müller-Schärer H. 1992. Physiological and growth responses of Centaurea maculosa (Asteraceae) to root herbivory under varying levels of interspecific plant competition and soil nitrogen availability. Oecologia 91:141–149.
- te Beest M, Le Roux JJ, Richardson DM, Brysting AK, Suda J, Kubesová M, Pysek P. 2012. The more the better? The role of polyploidy in facilitating plant invasions. Annals of Botany **109**:19–45.

- terHorst CP, Lau JA. 2015. Genetic variation in invasive species response to direct and indirect species interactions. Biological Invasions 17:651–659.
- Throop HL, Lerdau MT. 2004. Effects of nitrogen deposition on insect herbivory: implications for community and ecosystem processes. *Ecosystems* 7:109–133.
- Tilman D. 1989. Competition, nutrient reduction and the competitive neighbourhood of a bunchgrass. Functional Ecology 3:215–219.
- Turlings TCJ, Erb M. 2018. Tritrophic interactions mediated by herbivoreinduced plant volatiles: mechanisms, ecological relevance, and application potential. Annual Review of Entomology 63:433–452.
- Turner KG, Fréville H, Rieseberg LH. 2015. Adaptive plasticity and niche expansion in an invasive thistle. Ecology and Evolution 5:3183–3197.
- Uyi OO, Zachariades C, Hill MP. 2016. Nitrogen fertilisation improves growth of Chromolaena odorata (Asteraceae) and the performance of the biological control agent, Pareuchaetes insulata (Erebidae). Biocontrol Science and Technology 26:373–385.
- Valladares F, Sanchez-Gomez D, Zavala MA. 2006. Quantitative estimation of phenotypic plasticity: bridging the gap between the evolutionary concept and its ecological applications. *Journal* of Ecology 94:1103–1116.
- Vasquez E, Sheley R, Svejcar T. 2008. Nitrogen enhances the competitive ability of cheatgrass (Bromus tectorum) relative to native grasses. *Invasive Plant Science and Management* 1:287–295.
- Wang T, Hu J, Liu C, Yu D. 2017. Soil type can determine invasion success of Eichhornia crassipes. Hydrobiologia 788:281–291.
- Ward SM, Gaskin JF, Wilson LM. 2008. Ecological genetics of plant invasion: what do we know? Invasive Plant Science and Management 1:98–109.
- Wheeler G, Center T. 1996. The influence of hydrilla leaf quality on larval growth and development of the biological control agent Hydrellia pakistanae (Diptera: Ephydridae). Biological Control **7**:1–9.
- Wheeler GS, Center TD. 1997. Growth and development of the biological control agent Bagous hydrillae as influenced by hydrilla (Hydrilla verticillata) stem quality. Biological Control 8:52–57.
- Williams WI, Friedman JM, Gaskin JF, Norton AP. 2014. Hybridization of an invasive shrub affects tolerance and resistance to defoliation by a biological control agent. *Evolutionary Applications* 7:381–393.
- Williams DA, Overholt WA, Cuda JP, Hughes CR. 2005. Chloroplast and microsatellite DNA diversities reveal the introduction history of Brazilian peppertree (Schinus terebinthifolius) in Florida. Molecular Ecology 14:3643–3656.
- Wilson JRU, Rees M, Ajuonu O. 2007. Population regulation of a classical biological control agent larval density dependence in Neochetina eichhorniae (Coleoptera: Curculionidae), a biological control agent of water hyacinth Eichhornia crassipes. Bulletin of Entomological Research 96:145–152.
- Woo I, Zedler JB. 2002. Can nutrients alone shift a sedge meadow towards dominance by the invasive Typha × glauca? Wetlands 22:509–521.
- Wright P, Cregger MA, Souza L, Sanders NJ, Classen AT. 2014. The effects of insects, nutrients, and plant invasion on community structure and function above-and belowground. *Ecology and Evolution* 4:732–742.
- Xu CY, Julien MH, Fatemi M, Girod C, Van Klinken RD, Gross CL, Novak SJ. 2010. Phenotypic divergence during the invasion of Phyla canescens in Australia and France: evidence for selection-driven evolution. Ecology Letters 13:32–44.
- Yoshizuka EM, Roach DA. 2011. Plastic growth responses to simulated herbivory. International Journal of Plant Sciences 172:521–529.
- Younginger BS, Sirová D, Cruzan MB, Ballhorn DJ. 2017. Is biomass a reliable estimate of plant fitness? Applications in Plant Sciences 5:1600094.
- Yu H, Wang L, Liu C, Fan S. 2018. Coverage of native plants is key factor influencing the invasibility of freshwater ecosystems by exotic plants in China. Frontiers in Plant Science **9**:1–9.
- Zhang Z, Pan X, Blumenthal D, van Kleunen M, Liu M, Li B. 2018. Contrasting effects of specialist and generalist herbivores on resistance evolution in invasive plants. *Ecology* **99**:866–875.